

# VULNERABILITY OF THE ASIAN TYPICAL STEPPE TO GRAZING AND CLIMATE CHANGE

LINDSEY CHRISTENSEN<sup>1</sup>, MICHAEL B. COUGHENOUR<sup>2</sup>, JAMES E. ELLIS<sup>2</sup> and ZUO ZHONG CHEN<sup>3</sup>

<sup>1</sup>*Center for Environmental Science and Policy, Stanford University, Stanford, CA 94305-6055, U.S.A.*

*E-mail: lindsc@stanford.edu*

<sup>2</sup>*Natural Resource Ecology Laboratory, Colorado State University, Fort Collins, CO 80523, U.S.A.*

*E-mail: mikec@nrel.colostate.edu*

<sup>3</sup>*Ecology Research Center, Institute of Botany, Chinese Academy of Science, Beijing, China 100093*

*E-mail: chenzz@ns.ibcas.ac.cn*

**Abstract.** The vulnerability of grassland vegetation in Inner Mongolia to climate change and grazing was examined using an ecosystem model. Grazing is an important form of land use in this region, yet there are uncertainties as to how it will be affected by climate change. A sensitivity analysis was conducted to study the effects of increased minimum and maximum temperatures, ambient and elevated CO<sub>2</sub>, increased or decreased precipitation, and grazing on vegetation production. Simulations showed that herbaceous above ground net primary production was most sensitive to changes in precipitation levels. Combinations of increased precipitation, temperature, and CO<sub>2</sub> had synergistic effects on herbaceous production, however drastic increases in these climate scenarios left the system vulnerable to shifts from herbaceous to shrub-dominated vegetation when grazed. Reduced precipitation had a negative effect on vegetation growth rates, thus herbaceous growth was not sustainable with moderate grazing. Shifts in temporal biomass patterns due to changed climate have potentially significant implications for grazing management, which will need to be altered under changing climate to maintain system stability.

## 1. Introduction

Understanding the vulnerability of ecosystems to global change is becoming increasingly important as the effects of climate change are surfacing. It is well recognized that atmospheric changes have occurred worldwide, including increased CO<sub>2</sub> (Keeling et al., 1995; IPCC, 1996; Watson et al., 1996; Diaz et al., 1998; IPCC, 2001), increased temperatures (Karl et al., 1993; IPCC, 1996; Smit and Yunlong, 1996; IPCC, 2001), and changed precipitation patterns (Fu and Wen, 1999; Smit and Yunlong, 1996; Zhai et al., 1999; IPCC, 2001). Although large-scale climate changes have occurred, changes are predicted to differ depending on regional location. For example, increased winter temperatures and decreased precipitation patterns are only predicted in central and northern China (Smit and Yunlong, 1996; Polley et al., 2000).



*Climatic Change* **63**: 351–368, 2004.

© 2004 Kluwer Academic Publishers. Printed in the Netherlands.

Multiple studies have been conducted to examine the effects of localized climate changes on vegetation processes, and the scientific community recognizes the importance of adding grazing to this equation (Baker et al., 1993; Thornley and Cannell, 1997; Diaz et al., 1998; Hall et al., 1998; Reido et al., 2000; Ludwig et al., 2001). A significant obstacle to assessing grazing effects is the difficulty of conducting experiments at a large enough scale. Experiments have been conducted in chambers or sample plots to examine the interactive effects of climate change and grazing by insects or by clipping (Newton et al., 1995; Diaz et al., 1998), but few studies have explored the interactions of climate change and livestock grazing at larger scales (Baker et al., 1993; Panario, 1997). Grazing plays a significant role in ecosystem function (Watson et al., 2000), therefore it is important to include grazing in studies of ecosystem responses to climate change (Polley et al., 2000).

Native grasses of the typical steppe, defined as a bunch-grass steppe with many or few forbs in a semi-arid climate (Lavrenko and Karamysheva, 1993; Zhu, 1993), are well adapted to herbivory and have supported grazing as the dominant land use for thousands of years. A change in climate could affect this tightly coupled system, disrupt the balance between grazers and vegetation, and have negative effects on the people who use them (Alward et al., 1999). For these reasons, it is important to better understand interactions between climate change and grazing, to prevent overgrazing and to maintain viable grazing systems. To best analyze climate-grazing interactions, it is vital to simultaneously consider ecosystem components, such as vegetation, soil, and livestock (Allen-Diaz, 1996).

The objective of this study was to assess potential effects of climate change and grazing on the typical steppe grasslands using the SAVANNA ecosystem model (Coughenour, 1993). A factorial experiment involving precipitation, temperature, and CO<sub>2</sub> was conducted to explore grazing system sustainability under climatic change scenarios.

## 2. Methods

### 2.1. STUDY AREA

The typical steppe region of Inner Mongolia, China extends across 41° to 47° north latitude and 109° to 117° east longitude. This study was focused on the Baiinxile Livestock Farm located at 43.5° north latitude and 116.5° east longitude within the typical steppe. The farm is 3,680 km<sup>2</sup>, divided into an administrative village and 12 branches. A branch is an administrative division of land, ranging in size from 79 to 655 km<sup>2</sup>. This landscape consists of large rolling hills, with elevation ranging from 963 to 1569 m. The typical steppe is dominated by C<sub>3</sub> grasses, which include *Leymus chinensis* and *Stipa grandis* (Li, 1978). Important forb and shrub species include *Artemisia frigida* and *Caragana microphylla* (Zhu, 1993). The grasses in this region have coexisted with native and domestic grazers for thousands of years and are therefore likely to be well adapted to grazing.

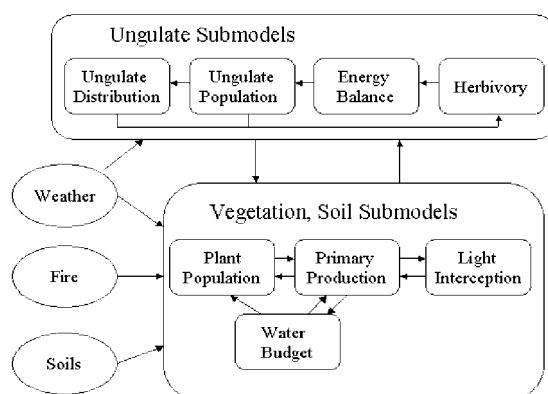


Figure 1. SAVANNA model structure (Coughenour, 1993).

Mean annual temperature is  $-0.4^{\circ}\text{C}$  with extremes ranging from  $-27.0$  to  $28.6^{\circ}\text{C}$  (Li, 1989; Xiao et al., 1997). Long term mean annual precipitation is 360 mm but varies between 180 and 500 mm. Because the majority of precipitation occurs during the growing season, biomass production is high as compared to other regions with similar amounts of rainfall but with smaller proportions received during the growing season.

## 2.2. MODEL DESCRIPTION

The SAVANNA model (Coughenour, 1993) used in this analysis is a spatially explicit, dynamic, simulation model that represents ecosystem processes at both local and regional scales. Ecosystem processes are modeled by using data on soils, topography, climate, and livestock densities to affect vegetation production and cover, livestock forage utilization, livestock production, and human offtake (Figure 1). SAVANNA is process-based, representing flows of biomass, nitrogen, and organisms, and subsequent rates of change in system states. Its one-week time step enables large-scale spatial and long-term temporal simulations. These capabilities enable simulations of interactive responses to climate and grazing.

SAVANNA has successfully been parameterized and used to simulate vegetation and ungulate production in a wide variety of ecosystems such as Rocky Mountain National Park, Colorado (Weisberg and Coughenour, 2003); Yellowstone National Park, Wyoming (Coughenour and Singer, 1996a,b); Kajiado District, Kenya (Boone et al., 2001); Northern Australia (Ludwig et al., 2001); Ngorongoro Conservation Area, Tanzania (Boone et al., 2002); Kruger National Park, South Africa (Kiker, 1998) and Inner Mongolia, China (Christensen et al., 2003).

A detailed description of the SAVANNA parameterization and validation for this study site can be found Christensen et al. (2003). Fieldwork was used to parameterize the vegetation and soil portions of the model (Xiao et al., 1996; Christensen et al., 1998) including vegetation biomass and percent cover data.

Information for livestock parameters, including height, weight, and diet were gathered from data specific to the Typical steppe (weight of adult sheep, horse, and cow = 75, 386, and 600 kg respectively) (Yang, 1987). Herbaceous plants were selected as the preferred forage type due to the overall dominance of grass and forbs in this ecosystem (personal communications with herders). Predicted green leaf biomass amounts were validated using 11 years of Normalized Difference Vegetation Index bimonthly composites derived from satellite imagery (Tucker, 1979; Justice et al., 1985; Malingreau, 1986). Correlations between observed and predicted values ranged from  $R^2 = 0.34$  to  $R^2 = 0.57$  ( $P < 0.001$ ) (Christensen et al., 2003).

The model was modified to represent  $\text{CO}_2$  effects on photosynthesis and transpiration. Under doubled  $\text{CO}_2$  (700 ppm) stomatal conductance and transpiration rates were reduced by 20%, while photosynthesis rates were increased by 20%. Conservative rather than maximum values were chosen based on field information (Jackson et al., 1994, 1998; Wand et al., 1999; J. Morgan, personal communication) because the modeled  $\text{CO}_2$  effect was constant, and not dependent on time of day or season. The  $\text{CO}_2$  response increased water use efficiency (*WUE*) and soil water, which could further affect plant growth (Knapp et al., 1996). The Penman–Monteith equation (Penman, 1953; Monteith, 1965) calculates canopy-scale transpiration, driven by temperature, radiation, and humidity. This formulation represents plant physiologically-mediated response to climate change by using temperature, radiation, and vapor pressure deficit as well as stomatal conductance to predict water loss from leaves. In the Penman–Monteith formula, an increase in temperature increases vapor pressure deficit, which results in an increase in transpiration rate. Stomatal conductance is differentially affected by temperature and relative humidity following Ball (1988). The interactions between climate,  $\text{CO}_2$ , and grazing are non-linear. Plant growth is an outcome of photosynthetic rate per unit leaf area, multiplied by leaf area index (LAI). It is non-linear response due to this multiplicative interaction. The effects of grazing on LAI, and the effects of  $\text{CO}_2$  on photosynthetic rates, can therefore interact non-linearly.

### 2.3. EXPERIMENTAL DESIGN

Three simulation experiments were conducted to examine temperature, precipitation,  $\text{CO}_2$ , and grazing effects on grassland vegetation. The first two experiments examined the sustainability of a grassland system impacted with different climate change scenarios (*CC*) (Riedo et al., 1997) including increased temperature, precipitation changes, and moderate grazing. Moderate grazing was defined as 39 animal units (*AU*)  $\text{km}^{-2}$ . This value was based on previous results (Christensen et al., 2003) as the maximum number of animals that could be supported by herbaceous vegetation in this region while maintaining a sustainable system of grazing and herbaceous biomass production. Higher animal densities resulted in degradation. The model averages animal units over the entire simulated area, so at any particular

moment, some areas could have a higher number of specified animals, whereas others could have fewer.

Experiment one was a sensitivity analysis, which used a factorial design with two temperature levels, three precipitation levels, and moderate grazing. Minimum and maximum temperatures were increased 2 and 0 °C respectively ( $T_{+2+0}$ ), then 5 and 1 °C ( $T_{+5+1}$ ) respectively. Precipitation levels included 80, 100 and 120% of actual precipitation data ( $P_{80}$ ,  $P_{100}$ ,  $P_{120}$ ). Experiment two was a factorial design, which followed the same design as experiment one, but with doubled CO<sub>2</sub> (700 ppm ( $C_{2x}$ )).

A third experiment was conducted where moderate grazing was replaced with light grazing from the first two experiments. The purpose was to examine the importance of grazing management when the combination of moderate grazing and climate change resulted in vegetation decline. Light grazing was defined in a previous study (Christensen et al., 2003) as 18 AU km<sup>-2</sup>, the maximum number of animals per km<sup>-2</sup> (within a range of 6 to 18 AU) that resulted in only slight decreases in herbaceous production.

Each experiment consisted of model runs 50 years in length. Eighteen years of local climate data were available from the area being simulated. To create 50 years of data, years were selected randomly from the 18 to create 32 more years (Figure 2). For climate change scenarios, precipitation days and temperatures were altered uniformly within each year, i.e., climate was not varied intra-annually. Thus each monthly input of temperature was increased and each monthly input of precipitation data was increased or decreased by 20%. The output variables used in the analyses were herbaceous and shrub above ground net primary production ( $ANPP_h$  and  $ANPP_s$ ).

### 3. Modeling Results

#### 3.1. EXPERIMENT 1

Combinations of ambient and increased temperatures ( $t_a$ ,  $T_{+2+0}$ ), ambient and increased precipitation ( $P_{100}$ ,  $P_{120}$ ), and moderate grazing had varying effects on  $ANPP_h$ , but these combinations maintained a sustainable level of herbaceous biomass (Table I). A  $T_{+2+0}$  increase in temperature increased  $ANPP_h$  by 3%, from 154 g m<sup>-2</sup> y<sup>-1</sup> with ambient temperature and precipitation ( $T_a P_{100}$ ) to 158 g m<sup>-2</sup> y<sup>-1</sup> (values are averaged for the 50 years simulation run). Adding a 20% increase in precipitation ( $T_{+2+0} P_{120}$ ) resulted in a 38% increase in  $ANPP_h$  compared to a 39% increase with  $T_a P_{120}$ .

Climate change scenarios (CC) with a 20% reduction in precipitation, ambient and elevated temperatures ( $T_a$ ,  $T_{+2+0}$ ), and moderate grazing resulted in herbaceous productivity initially following precipitation patterns in the first 12 years, but in an overall decrease in herbaceous productivity. The CC of ambient temperatures at 80% precipitation ( $T_a P_{80}$ ) and increased temperature at 80% precipitation

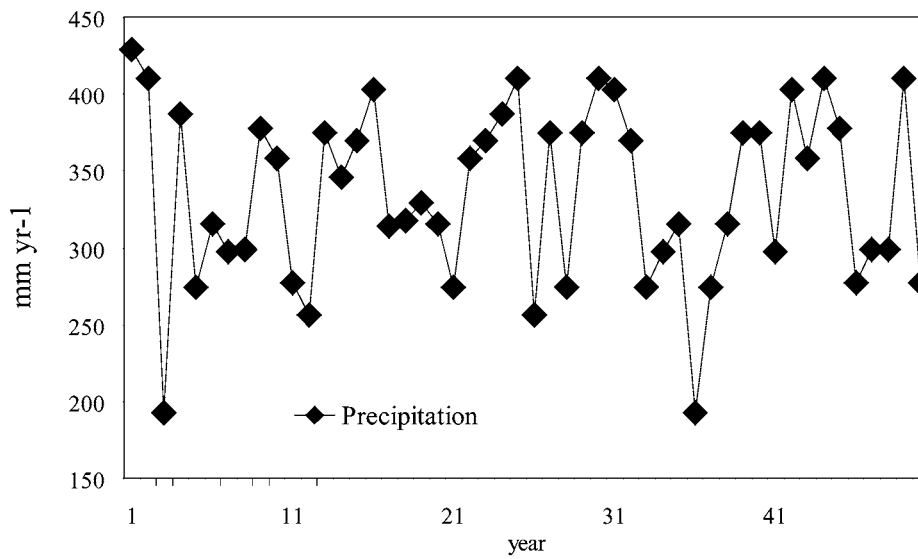


Figure 2. Annual precipitation (mm) for 50-year simulation run.

Table I

Herbaceous above ground net primary production ( $\text{g m}^{-2} \text{yr}^{-1}$ ) responses to ambient and elevated  $\text{CO}_2$  ( $\mu\text{mol mol}^{-1}$ ), increased temperature ( $^{\circ}\text{C}$ ), precipitation (*Ppt*) change (% of ambient), and moderate grazing ( $29 \text{ AU km}^{-2}$ ) from experiments 1 and 2. Values are averaged over 50-year simulation.

<i>Ppt</i> (%)	Temperature ( $^{\circ}\text{C}$ )		350 $\mu\text{mol mol}^{-1}$ $\text{CO}_2$ (Exp. 1)	700 $\mu\text{mol mol}^{-1}$ $\text{CO}_2$ (Exp. 2)
	$\Delta T_{\text{min}}$	$\Delta T_{\text{max}}$		
<b>80</b>	<b>0</b>	<b>0</b>	34.6	130.7
	<b>2</b>	<b>0</b>	29.7	120.5
	<b>5</b>	<b>1</b>	22.3	34.4
<b>100</b>	<b>0</b>	<b>0</b>	<b>153.9</b>	186.9
	<b>2</b>	<b>0</b>	158.2	187.4
	<b>5</b>	<b>1</b>	30.8	183.3
<b>120</b>	<b>0</b>	<b>0</b>	214.0	239.4
	<b>2</b>	<b>0</b>	212.6	240.3
	<b>5</b>	<b>1</b>	142.0	211.2

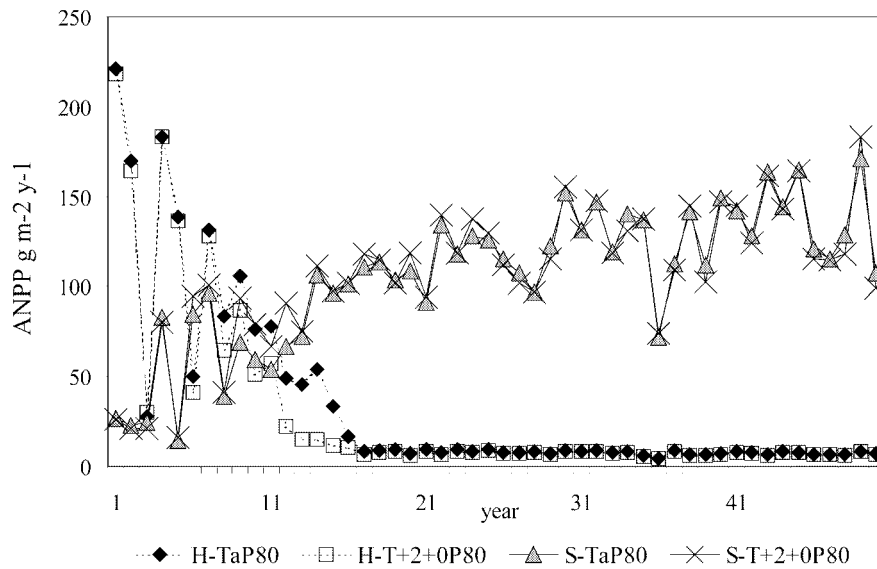


Figure 3. Herbaceous (*H*) and shrub (*S*) above ground net primary production from 50-year simulation runs with ambient and increased temperature ( $T_a$ ,  $T_{+2+0}$ ), and 80% precipitation ( $P_{80}$ ), and moderate grazing.

( $T_{+2+0}P_{80}$ ) reduced  $ANPP_h$  by an average of 78 and 81% respectively. Shrub above ground net primary production ( $ANPP_s$ ) increased in response to decreased  $ANPP_h$ . Shrubby vegetation was not browsed, so it expanded as  $ANPP_h$  declined and competition with herbs decreased (Figure 3).

Combinations of  $T_{+5+1}$ , all precipitation levels, and moderate levels of grazing had greater impacts on  $ANPP_h$  than the smaller temperature increase. A large temperature increase  $T_{+5+1}$  ( $T_{+5+1}P_{100}$ ) in combination with moderate grazing reduced  $ANPP_h$  catastrophically to an average of  $30.8 \text{ g m}^{-2} \text{ y}^{-1}$  (Table I). When precipitation was increased by 20% ( $T_{+5+1}P_{120}$ ),  $ANPP_h$  still decreased but much less, from  $153.9 \text{ g m}^{-2} \text{ y}^{-1}$  to  $142.0 \text{ g m}^{-2} \text{ y}^{-1}$ . Reduction in precipitation ( $CC$  of  $T_{+5+1}P_{80}$ ) decreased  $ANPP_h$  by an average of 86% from  $T_aP_{100}$ . There was a rapid increase of shrubs due to shrub access to water in deep soil layers and lack of browsing of the shrub vegetation (Figure 4). The dramatic increase in shrub production with increased temperature and precipitation ( $T_{+5+1}P_{120}$ ) occurred after a two year period with below average precipitation. Shrubs were able to capitalize on available resources and gain a competitive advantage over herbaceous production grazed with moderate levels of grazing during these low precipitation years.

### 3.2. EXPERIMENT 2

Elevated  $\text{CO}_2$  with ambient precipitation, all temperature levels, and moderate grazing resulted in increases in  $ANPP_h$ . Both herbaceous and shrub production increased with doubled  $\text{CO}_2$ , due to increased water use efficiency and soil water

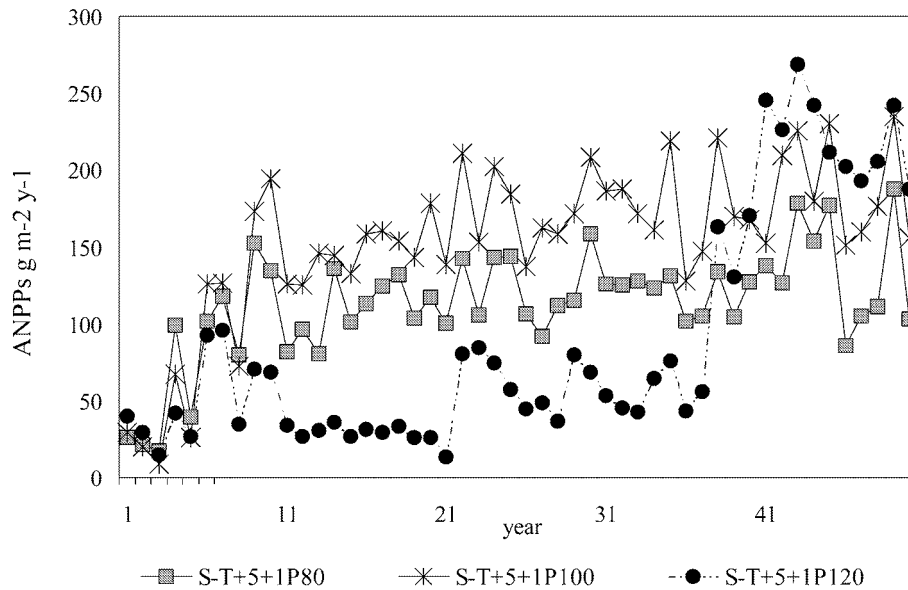


Figure 4. Shrub (*S*) above ground net primary production from 50-year simulation runs with increased temperature ( $T_{+5+1}$ ), 80, 100 and 120% ambient precipitation ( $P_{80}$ ,  $P_{100}$ ,  $P_{120}$ ), and moderate grazing.

(Figure 5). The  $T_a P_{100} C_{2x}$ ,  $T_{+2+0} P_{100} C_{2x}$ , and  $T_{+5+1} P_{100} C_{2x}$  scenarios resulted in 21, 22 and 19% increases in  $ANPP_h$  respectively (Table I). When precipitation was increased by 20%, there were much larger increases in  $ANPP_h$ .  $CC$  of  $T_a P_{120} C_{2x}$ ,  $T_{+2+0} P_{120} C_{2x}$ , and  $T_{+5+1} P_{120} C_{2x}$  resulted in 56, 56 and 37% increases in  $ANPP_h$  respectively. A doubling of  $CO_2$  did not ameliorate the effect of a 20% decrease in precipitation and moderate grazing on plant production, and  $ANPP_h$  decreased in all  $CC$  scenarios. The  $T_{+5+1} P_{80} C_{2x}$  scenario resulted in a 78% decrease in  $ANPP_h$ .  $CC$  of  $T_a P_{80} C_{2x}$  and  $T_{+2+0} P_{80} C_{2x}$  resulted in a gradual decrease in  $ANPP_h$  but no shift in vegetation composition (Figure 6).

### 3.3. EXPERIMENT 3

A decrease in grazer density reversed declines in  $ANPP_h$  and maintained a sustainable level of herbaceous biomass production. When grazing was reduced from moderate to light with  $CC$  of  $T_{+5+1} P_{120}$ , average  $ANPP_h$  increased to 228.2  $g\ m^{-2}\ y^{-1}$  as compared to 142.0  $g\ m^{-2}\ y^{-1}$  with a moderate grazing density (Table II). The  $T_{+5+1} P_{100}$  scenario and light grazing was also sustainable, with an average of 170.1  $g\ m^{-2}\ y^{-1}$  (Figure 7). Light grazing and  $T_a P_{80} C_{2x}$ ,  $T_{+2+0} P_{80} C_{2x}$ ,  $T_{+5+1} P_{80} C_{2x}$ ,  $T_a P_{80}$ , and  $T_{+2+0} P_{80}$  scenarios were sustainable for the 50-year simulation run with an average  $ANPP_h$  of 151.3  $g\ m^{-2}\ y^{-1}$ , 149.4  $g\ m^{-2}\ y^{-1}$ , 135.9  $g\ m^{-2}\ y^{-1}$ , 129.7  $g\ m^{-2}\ y^{-1}$ , and 117.6  $g\ m^{-2}\ y^{-1}$  respectively, but  $ANPP_h$  values gradually decreased over the simulation period. Light grazing in combination with



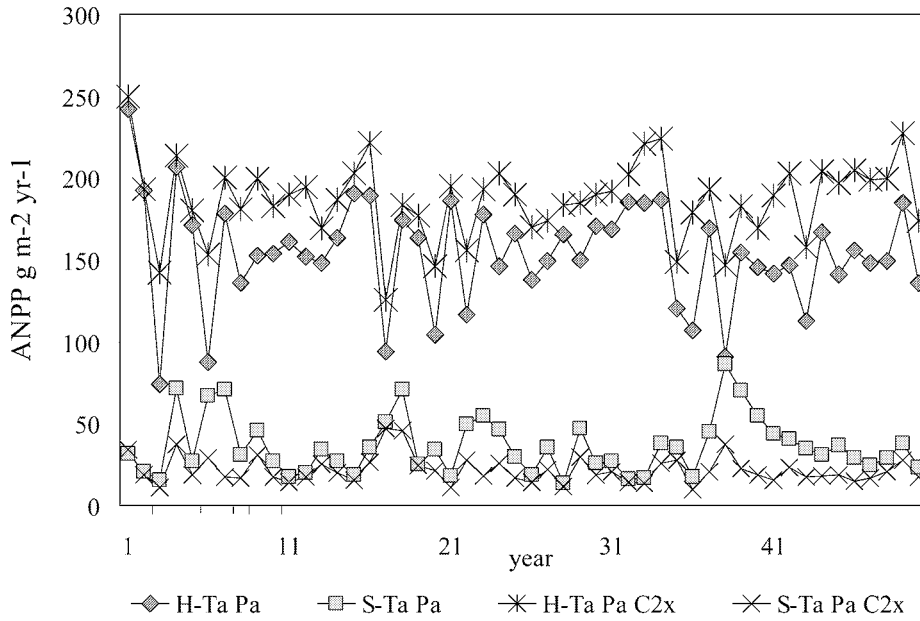


Figure 5. Herbaceous (*H*) and shrub (*S*) above ground net primary production with ambient and elevated CO<sub>2</sub> from 50-year simulation runs with ambient temperature (*T*) and precipitation (*P*) and moderate grazing.

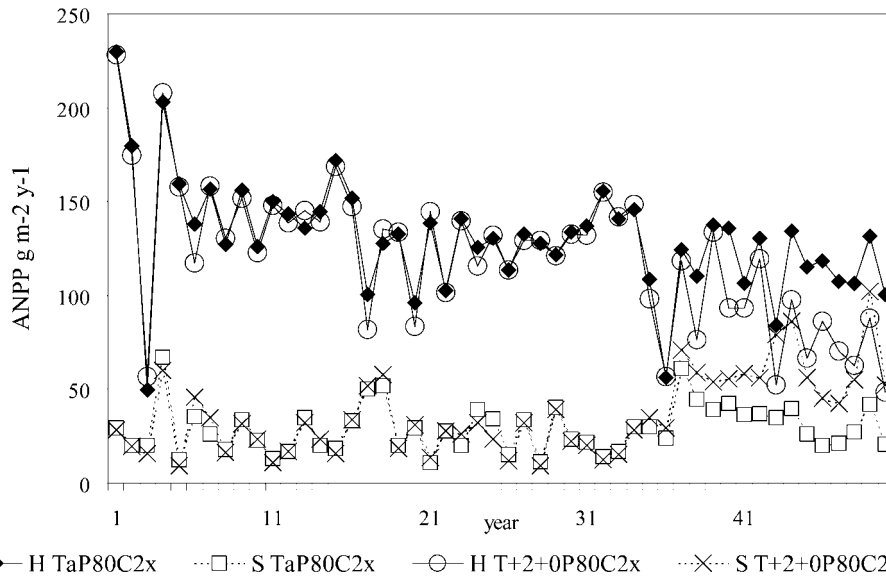


Figure 6. Herbaceous (*H*) and shrub (*S*) ANPP from 50-year simulation runs with ambient and increased temperature (*T<sub>a</sub>*, *T<sub>+2+0</sub>*), 80% precipitation (*P<sub>80</sub>*), elevated CO<sub>2</sub> (*C<sub>2x</sub>*), and moderate grazing.

Table II

Herbaceous above ground net primary production ( $\text{g m}^{-2} \text{ yr}^{-1}$ ) responses to ambient and elevated  $\text{CO}_2$  ( $\mu\text{mol mol}^{-1}$ ), increased temperature ( $^{\circ}\text{C}$ ), precipitation ( $Ppt$ ) change (% of ambient), and light grazing ( $18 \text{ AU km}^{-2}$ ) from experiment 3. Values are averaged over 50-year simulation.

$Ppt$ (%)	Temperature ( $^{\circ}\text{C}$ )		350	700
	$\Delta T_{\text{min}}$	$\Delta T_{\text{max}}$	$\mu\text{mol mol}^{-1}$ $\text{CO}_2$ (Exp. 1)	$\mu\text{mol mol}^{-1}$ $\text{CO}_2$ (Exp. 2)
<b>80</b>	<b>0</b>	<b>0</b>	129.7	151.3
	<b>2</b>	<b>0</b>	117.6	149.4
	<b>5</b>	<b>1</b>	42.1	135.9
<b>100</b>	<b>0</b>	<b>0</b>	–	–
	<b>2</b>	<b>0</b>	–	–
	<b>5</b>	<b>1</b>	170.1	–
<b>120</b>	<b>0</b>	<b>0</b>	–	–
	<b>2</b>	<b>0</b>	–	–
	<b>5</b>	<b>1</b>	228.2	–

$CC T_{+5+1} P_{80}$  was not sustainable and  $ANPP_h$  production was reduced to  $42.1 \text{ g m}^{-2} \text{ y}^{-1}$ . The system was unable to maintain herbaceous growth with decreased precipitation despite the reduction in grazing density.

#### 4. Discussion

Simulations showed that  $ANPP_h$  in this grazing ecosystem was most sensitive to changes in precipitation levels. However combinations of precipitation, temperature, and  $\text{CO}_2$  had synergistic effects on herbaceous production. Increases in precipitation had larger positive effects on  $ANPP_h$  than a doubling of  $\text{CO}_2$ , while small increases in temperature only had minimal effects. Greater increases in temperature put a larger stress on the plants, resulting in declining  $ANPP_h$  unless increased precipitation,  $\text{CO}_2$ , or the combination of the two ameliorated the effects of the large increase in temperature. Because net primary production in this region is water limited, sufficient soil moisture was necessary for grasses to recover from grazing.

Reduced precipitation caused a decline in  $ANPP_h$ , so that moderate grazing of  $39 \text{ AU km}^{-2}$  could not be sustained in this modeling experiment. A decrease

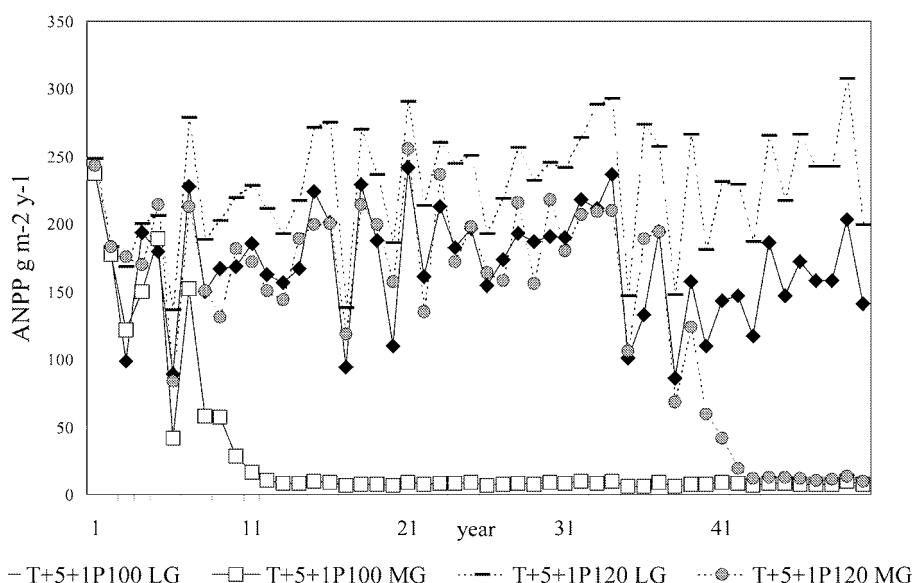


Figure 7. Herbaceous above ground net primary production ( $ANPP_h$ ) in response to increased temperature ( $T_{+5+1}$ ), and precipitation ( $P_{100}$ ,  $P_{120}$ ) change (% of ambient) with light grazing ( $LG = 18 AU km^{-2}$ ) for 50-year simulation runs.  $ANPP_h$  also reported for these climate change scenarios with moderate grazing (MG).

in precipitation resulted in lower soil water, which decreased plant growth rates. Lower growth rates were not sufficient for herbaceous plants to respond from grazing and shrubs were able to capitalize on available resources once the herbaceous vegetation was suppressed (Archer, 1996). Because shrubs were not browsed they displaced the overgrazed herbaceous plants and became the dominant functional group. Doubling  $CO_2$  partially ameliorated the effects of reduced precipitation as a result of increased  $WUE$ . A doubling of  $CO_2$  caused a decrease in stomatal conductance and increased photosynthesis; therefore  $WUE$  increased, which in turn increased plant growth in this water-limited system. The  $CO_2$  effects allowed grazed grasses to remain competitive relative to unbrowsed shrubs despite a decrease in water.

Modeling experiments showed the system could not maintain a 50-year sustainable level of herbaceous biomass when subject to certain grazing and climate patterns; thus it was vulnerable to a shift from herbaceous to shrub-dominated vegetation. Shrub growth was limited because grasses limit water availability to shrubs with their intensive fibrous root structure (Köchy and Wilson, 2000; Polley et al., 2000; Brown et al., 1998). Decreased precipitation in combination with grass removal from grazing and lack of shrub browsing allowed shrub encroachment. It is important to note this system was dominated by grazers, including sheep, horses, and cattle. If goats, which browse shrubs, were simulated then the results would be different. These vegetation changes due to grazing and climate change are similar

to threshold state changes typical of non-linear systems. It is important to recognize the potentials for vegetation shifts with climate change because they can lead to permanent and sudden ecological changes with little forewarning. Managers need to monitor livestock offtake to reduce system vulnerability under such scenarios of climatic change.

Previous models have projected increased grassland production under elevated CO<sub>2</sub> (Parton et al., 1995; Coughenour and Chen, 1997; Gao and Zhang, 1997; Neilson et al., 1998) although some of these models did not incorporate direct CO<sub>2</sub> effects on photosynthesis. Coughenour and Parton (1997) used the mechanistic GRASS-CSOM model and found elevated CO<sub>2</sub> at ambient and increased precipitation levels resulted in a 20–70% increase in C<sub>3</sub> grassland production. Wand et al. (1999) in a meta-analysis concluded grass biomass on average increased by 44% with elevated CO<sub>2</sub>. While biomass could not be directly related to production, the trends in both show increases. A simulation study conducted by Gao and Zhang (1997) predicted an overall 15% increase in vegetation production with doubled CO<sub>2</sub> in the northeast region of China. However, responses of specific sites to elevated CO<sub>2</sub> differed with 26 and 61% increases in net primary production (NPP) in *Stipa baicalensis* steppe and *Aneurolipidium chinense* steppe grasslands, respectively. The combination of doubled CO<sub>2</sub> and increased precipitation and temperature caused a 24% decrease in vegetation production in the *Aneurolipidium chinense* steppe while it increased NPP by 44% in the *Stipa baicalensis* steppe. Decreases were due to larger temperature effects on the *Aneurolipidium* steppe than the *Stipa* steppe (Gao and Yu, 1998). SAVANNA was not parameterized to simulate differences in these species to elevated CO<sub>2</sub> or temperature.

This modeling experiment explored the possible implications of temperature, precipitation, and CO<sub>2</sub> changes for grazing management of the grasslands in North-east Asia. A small increase in temperature, with increased CO<sub>2</sub> and decreased precipitation is likely to occur in this region within the next hundred years (Keeling et al., 1995; Smit and Yunlong, 1996; Giorgi et al., 1998; Polley et al., 2000; Watson et al., 2000; IPCC, 2001). The simulation model estimates a 22% decrease in ANPP<sub>h</sub> under such a climate regime and moderate grazing. To maintain ecosystem states it would be necessary to decrease livestock numbers in proportion to the decrease in forage production (Coughenour and Chen, 1997).

A second implication is the possible shift of vegetative state due to climatic change and grazing. Many rangelands exist in areas with variable climate patterns. Therefore, small shifts in climate could increase system vulnerability to vegetation state change (Rietkerk and van de Koppel, 1997; Ludwig et al., 2001). Rietkerk and van de Koppel (1997) give a mechanistic explanation of grazing induced shifts in vegetative stable states. They used a model to show how a loss in plant growth with heavy grazing resulted in a decrease in soil water due to higher run-off and decreased percolation. Lower soil water subsequently decreased plant production, so resources were insufficient to sustain production while being grazed. They found semi-arid systems are vulnerable to threshold effects and are thus fragile,

when resources have reached these low levels. Results here are consistent with that conclusion.

With expanding human populations, there are increased pressures on rangelands for their resources (Polley et al., 2000). The typical steppe region is particularly vulnerable because it is currently experiencing climate change and an increasing trend in human and livestock population densities coupled with increased sedentarization. This intensified form of land-use could have disastrous effects with climate change (Archer, 1995). Models have been used to address the effects of grazing and climate on land use processes (Parton et al., 1987; Hanson et al., 1988; Thornley and Verberne, 1989; Hunt et al., 1991; Coughenour and Chen, 1997; Thornley and Cannell, 1997; Hall et al., 1998; Riedo et al., 2000; Ludwig et al., 2001) but there remain many unanswered questions (Nosberger et al., 2000). Studies are only beginning to analyze how livestock based agriculture will change due to shifts in climate patterns (Parsons et al., 2001).

Model simulations from this analysis give detailed insight as to which climate change scenarios had the greatest impact on changes in biomass. Grassland ecosystems are vulnerable to climate change and will likely experience increases or decreases in plant production. But more importantly, climate change can destabilize grazing ecosystems if not managed properly. Systems that would otherwise remain unsustainable under increased temperatures and decreased precipitation can maintain their stability if livestock grazing densities are adjusted.

## 5. Memorial

Dr. James Ellis was unfortunately unable to see the completion of this work due to his untimely death in March of 2002. With this paper, we would like to honor the efforts Dr. Ellis made with his research accomplishments in the Inner Mongolia region and would like to recognize his preeminent work on understanding the integration of humans and natural processes in arid ecosystems globally. We will miss him greatly.

## Acknowledgements

The authors wish to thank the NREL, Colorado State University, Fort Collins, CO 80523 and the Inner Mongolia Grassland Ecosystem Research Station, Inner Mongolia, China. We gratefully acknowledge financial support from the National Science Foundation Grant OPP.I.no.95-9.

### Nomenclature

<i>CC</i>	climate change scenarios
$T_a P_{100}$	ambient temperature, ambient precipitation
$T_a P_{100} C_{2x}$	ambient temperature, ambient precipitation, doubled CO <sub>2</sub>
$T_a P_{80}$	ambient temperature, 80% precipitation
$T_a P_{80} C_{2x}$	ambient temperature, 80% precipitation, doubled CO <sub>2</sub>
$T_a P_{120}$	ambient temperature, 120% precipitation
$T_a P_{120} C_{2x}$	ambient temperature, 120% precipitation, doubled CO <sub>2</sub>
$T_{+2+0} P_{100}$	increase minimum temperature by 2 °C, max by 0 °C, ambient precipitation
$T_{+2+0} P_{100} C_{2x}$	increase minimum temperature by 2 °C, max by 0 °C, ambient precipitation, doubled CO <sub>2</sub>
$T_{+2+0} P_{80}$	increase minimum temperature by 2 °C, max by 0 °C, 80% precipitation
$T_{+2+0} P_{80} C_{2x}$	increase minimum temperature by 2 °C, max by 0 °C, 80% precipitation, doubled CO <sub>2</sub>
$T_{+2+0} P_{120}$	increase minimum temperature by 2 °C, max by 0 °C, 120% precipitation
$t_{+2+0} P_{120} C_{2x}$	increase minimum temperature by 2 °C, max by 0 °C, 120% precipitation, doubled CO <sub>2</sub>
$T_{+5+1} P_{100}$	increase minimum temperature by 5 °C, max by 1 °C, ambient precipitation
$T_{+5+1} P_{100} C_{2x}$	increase minimum temperature by 5 °C, max by 1 °C, ambient precipitation, doubled CO <sub>2</sub>
$T_{+5+1} P_{80}$	increase minimum temperature by 5 °C, max by 1 °C, 80% precipitation
$T_{+5+1} P_{80} C_{2x}$	increase minimum temperature by 5 °C, max by 1 °C, 120% precipitation, doubled CO <sub>2</sub>
$T_{+5+1} P_{120}$	increase minimum temperature by 5 °C, max by 1 °C, 120% precipitation
$T_{+5+1} P_{120} C_{2x}$	increase minimum temperature by 5 °C, max by 1 °C, 120% precipitation, doubled CO <sub>2</sub>

## References

- Allen-Diaz, B.: 1996, 'Rangelands in a Changing Climate: Impacts, Adaptations, and Mitigation', in Watson, R. T., Zinyowera, M. C., Moss, R. H., and Dokken, D. J. (eds.), *Climate Change 1995 – Impacts, Adaptations and Mitigation of Climate Change: Scientific-Technical Analyses*, Cambridge University Press, Cambridge, pp. 130–158.
- Alward, R. D., Detling, J. K., and Milchunas, D. G.: 1999, 'Grassland Vegetation Changes and Nocturnal Global Warming', *Science* **283**, 229–231.
- Archer, S.: 1996, 'Assessing and Interpreting Grass-Woody Plant Dynamics', in Hodgson, J. and Illius, A. W. (eds.), *The Ecology and Management of Grazing Systems*, CAB International, Oxford, pp. 101–134.
- Archer, S., Schimel, D., and Holland, E.: 1995, 'Mechanisms for Shrubland Expansion: Land Use, Climate or CO<sub>2</sub>', *Clim. Change* **29**, 91–99.
- Baker, B. B., Hanson, J. D., Bourdon, R. M., and Eckert, J. B.: 1993, 'The Potential Effects of Climate Change on Ecosystem Processes and Cattle Production on U.S. Rangelands', *Clim. Change* **25**, 97–117.
- Ball, J. T.: 1988, *An Analysis of Stomatal Conductance*, Doctor of Biological Sciences Thesis Dissertation, Stanford University, Stanford.
- Boone, R. B., Coughenour, M. B., Galvin, K. A., and Ellis, J. E.: 2001, 'Using the SAVANNA Modeling System to Address Potential Management Questions in Ngorongoro, Tanzania and Kajiado, Kenya', in Boone, R. B. and Coughenour, M. B. (eds.), *A System for Integrated Management and Assessment of East African Pastoral Lands: Balancing Food Security, Wildlife Conservation, and Ecosystem Integrity*, Report to the Global Livestock Collaborative Research Support Program, University of California, Davis, CA, U.S.A., pp. 73–104.
- Boone, R. B., Coughenour, M. B., Galvin, K. A., and Ellis, J. E.: 2002, 'Addressing Management Questions for Ngorongoro Conservation Area, Tanzania', *African Journal of Ecology*, in press.
- Brown, J. R., Scanla, J. C., and McIvor, J. G.: 1998, 'Competition by Herbs as a Limiting Factor in Shrub Invasion in Grassland: A Test with Different Growth Forms', *J. Vegetation Science* **9**, 829–836.
- Christensen, L., Coughenour, M. B., Ellis, J. E., and Chen, Z.: 2003, 'Sustainability of Inner Mongolian Grasslands: Application of the Savanna Model', *J. Range Manage.* **56**, 319–337.
- Christensen, L., Coughenour, M. B., Ellis, J., Li, L., and Han, Y.: 1998, 'Grazing Effects on the Typical Steppe Grassland of Inner Mongolia', *Proceedings of the LUTEA Conference*, Beijing, China.
- Coughenour, M. B.: 1993, SAVANNA – Landscape and Regional Ecosystem Model, User Manual, Colorado State University, Ft. Collins, CO, U.S.A.
- Coughenour, M. B. and Chen, D. X.: 1997, 'Assessment of Grassland Ecosystem Responses to Atmospheric Change Using Linked Plant-Soil Process Models', *Ecol. Appl.* **7**, 802–827.
- Coughenour, M. B. and Parton, W. J.: 1997, 'Integrated Models of Ecosystem Function: A Grassland Case Study', in Walker, B. and Steffen, W. (eds.), *Global Change and Terrestrial Ecosystems, International Geosphere-Biosphere Programme Book Series*, Cambridge University Press, Cambridge, pp. 93–114.
- Coughenour, M. B. and Singer, F. J.: 1996a, 'Yellowstone Elk Population Responses to Fire – A Comparison of Landscape Capacity and Spatial-Dynamic Ecosystem Modeling Approaches', in Greenlee, J. (ed.), *The Ecological Implications of Fire in Greater Yellowstone: Proceedings of the Second Biennial Conference on the Greater Yellowstone Ecosystem*, Yellowstone National Park, September 19–21, 1993, International Association of Wildland Fire, Fairfield, WA, pp. 169–180.
- Coughenour, M. B. and Singer, F. J.: 1996b, 'Elk Population Processes in Yellowstone National Park under the Policy of Natural Regulation', *Ecol. Appl.* **6**, 573–593.

- Diaz, S., Fraser, L. H., Grime, J. P., and Falczuk, V.: 1998, 'The Impact of Elevated CO<sub>2</sub> on Plant-Herbivore Interactions: Experimental Evidence of Moderating Effects at the Community Level', *Oecologia* **117**, 177–186.
- Fu, C. and Wen, G.: 1999, 'Variation of Ecosystems over East Asia in Association with Seasonal, Interannual and Decadal Monsoon Climate Variability', *Clim. Change* **43**, 477–494.
- Gao, Q. and Yu, M.: 1998, 'A Model of Regional Vegetation Dynamics and its Application to the Study of Northeast China Transect (NECT) Responses to Global Change', *Global Biogeochem. Cycles* **12**, 329–344.
- Gao, Z. and Zhang, X.: 1997, 'A Simulation Study of Responses of the Northeast China Transect to Elevated CO<sub>2</sub> and Climate Change', *Ecological Applications* **7**, 470–483.
- Giorgi, F., Meehl, G. A., Kattenberg, A., Grassl, H., Mitchell, J. F. B., Stouffer, Tokioka, T., Weaver, A. J., and Wigley, T. M. L.: 1998, 'Simulation of Regional Climate Change with Global Coupled Climate Models and Regional Modelling Techniques', in Watson, R. T., Zinyowera, M. C., Moss, R. H., and Dokken, D. J. (eds.), *The Regional Impacts of Climate Change: An Assessment of Vulnerability*, Cambridge University Press, New York, pp. 427–437.
- Hall, W. B., McKeon, G. M., Carter, J. O., Day, K. A., Howden, S. M., Scanlan, J. D., Johnston, P. W., and Burrows, W. H.: 1998, 'Climate Change in Queensland's Grazing Lands: II. An Assessment on the Impact on Animal Production from Native Pastures', *Rangeland Journal* **20**, 177–205.
- Hanson, J. D., Skiles, J. W., and Parton, W. J.: 1988, 'A Multi-Species Model for Rangeland Plant Communities', *Ecol. Modelling* **44**, 89–123.
- Hunt, H. W., Trlica, M. J., Redente, E. F., Moore, J. E., Detling, J. K., Kittel, T. G. F., Walter, D. E., Fowler, M. C., Klein, D. A., and Elliott, E. T.: 1991, 'Simulation Model for the Effects of Climate Change on Temperate Grassland Ecosystems', *Ecol. Modelling* **53**, 205–246.
- Intergovernmental Panel on Climate Change (IPCC): 1996, *Climate Change 1995: The Science of Climate Change. The Second IPCC Scientific Assessment*, Houghton, J. T., Meira Filho, L. G., Callendar, B. A., Harris, N., Kattenberg, A., and Maskell, K. (eds.), Cambridge University Press, N.Y., 572 pp.
- Intergovernmental Panel on Climate Change (IPCC): 2001, *Climate Change 2001: The Scientific Basis. Contribution of Working Group 1 to Third Assessment Report of the IPCC*, Houghton, J. and Ding, Y. (chairs), Cambridge University Press, Cambridge, 881 pp.
- Jackson, R. B., Sala, O. E., Field, E. B., and Mooney, H. A.: 1994, 'CO<sub>2</sub> Alters Water Use, Carbon Gain, and Yield for the Dominant Species in a Natural Grassland', *Oecologia* **98**, 257–262.
- Jackson, R. B., Sala, O. E., Paruelo, J. M., and Mooney, H. A.: 1998, 'Ecosystem Water Fluxes for Two Grasslands in Elevated CO<sub>2</sub>; A Modeling Analysis', *Oecologia* **113**, 537–546.
- Justice, C. O., Townshend, J. R. G., Holben, B. N., and Tucker, C. J.: 1985, 'Analysis of the Phenology of Global Vegetation Using Meteorological Satellite Data', *Int. J. Remote Sensing* **6**, 1271–1318.
- Karl, T. R., Jones, P. D., Knight, R. W., Kukla, G., Plummer, N., Razuvayev, V., Gallo, K. P., Lindsey, J., Charlson, R. J., and Peterson, T. C.: 1993, 'Asymmetric Trends of Daily Maximum and Minimum Temperature', *Bull. Amer. Meteorol. Soc.* **74**, 1007–1023.
- Keeling, C. D., Whorf, T. P., Wahlen, M., and van der Plicht, J.: 1995, 'Interannual Extremes in the Rate of Rise of Atmospheric Carbon Dioxide since 1980', *Nature* **375**, 666–670.
- Kiker, G. A.: 1998, *Development and Comparison of Savanna Ecosystem Models to Explore the Concept of Carrying Capacity*, Thesis Dissertation No. DAI, 59, No. 07B: 3569, Cornell University, Ithica, New York, 390 pp.
- Knapp, A. K., Hamerlynck, E. P., Ham, J. M., and Owensby, C. E.: 1996, 'Responses in Stomatal Conductance to Elevated CO<sub>2</sub> in 12 Grassland Species that Differ in Growth Form', *Vegetatio* **125**, 31–41.
- Köchy, M. and Wilson, S. D.: 2000, 'Competitive Effects of Shrubs and Grasses in Prairie', *OIKOS* **91**, 385–395.



- Lavrenko, E. M. and Karamysheva, Z. V.: 1993, 'Steppes of the Former Soviet Union and Mongolia', in Coupland, R. T. (ed.), *Ecosystems of the World. Natural Grasslands: Eastern Hemisphere and Resume*, Elsevier, New York, pp. 3–59.
- Li, J. D.: 1978, 'Aneurolepidium chinense Grassland in China', *J. Northeast Normal University* **1**, 145–159.
- Li, Y. H.: 1989, 'Impact of Grazing on Aneurolepidium Chinense Steppe and Stipa Grandis Steppe', *Acta Oecologica* **10**, 31–46.
- Ludwig, J. A., Coughenour, M. B., Liedloff, A. C., and Dyer, R.: 2001, 'Modelling the Resilience of Australian Savanna Systems to Grazing Impacts', *Environment International* **27**, 167–172.
- Malingreau, J. P.: 1986, 'Global Vegetation Dynamics: Satellite Observations over Asia', *Int. J. Remote Sensing* **7**, 1121–1146.
- Monteith, J. L.: 1965, 'Evaporation and Environment', *Symposia of the Society for Experimental Biology* **19**, 205–234.
- Neilson, R. P., Prentice, I. C., Smith, B., Kittel, T., and Viner, D.: 1998, 'Simulated Changes in Vegetation Distribution under Global Arming', in Watson, R. T., Zinyoweie, M. C., and Moss, R. H. (eds.), *The Regional Impacts of Climate Change: An Assessment of Vulnerability*, Cambridge University Press, N.Y., pp. 439–456.
- Newton, P. C. D., Clark, H., Bell, C. C., Glasgow, E. M., Tate, K. R., Ross, D. J., Yeates, G. W., and Sagar, S.: 1995, 'Plant Growth and Soil Processes in Temperate Grassland Communities at Elevated CO<sub>2</sub>', *J. Biogeogr.* **22**, 235–240.
- Nosberger, J., Blum, H., and Fuhrer, J.: 2000, 'Crop Ecosystem Responses to Climatic Change: Productive Grasslands', in Reddy, K. R. and Hodges, H. F. (eds.), *Climate Change and Global Crop Productivity*, CABI Publishing, N.Y., pp. 271–291.
- Panario, D. and Bidegain, M.: 1997, 'Climate Change Effects on Grasslands in Uruguay', *Clim. Res.* **9**, 37–40.
- Parsons, D. J., Armstrong, A. C., Turnpenny, J. R., Matthews, A. M., Cooper, K., and Clark, J. A.: 2001, 'Integrated Models of Livestock Systems for Climate Change Studies. 1. Grazing Systems', *Global Change Biology* **7**, 93–112.
- Parton, W. J., Schimel, D. S., Cole, C. V., and Ojima, D. S.: 1987, 'Analysis of Factors Controlling Soil Organic Matter Levels in Great Plains Grasslands', *Soil Science Society of America Journal* **51**, 1173–1197.
- Parton, W. J., Scurlock, J. M. O., Ojima, D. S., Schimel, D. S., Hall, D. O., and S. G. members: 1995, 'Impact of Climate Change on Grassland Production and Soil Carbon Worldwide', *Global Change Biology* **1**, 13–22.
- Penman, H. L.: 1953, 'The Physical Basis of Irrigation Control', *Report of the 13th International Horticultural Congress* **2**, 913–914.
- Polley, H. W., Morgan, J. A., Cambell, B. D., and Smith, M. S.: 2000, 'Crop Ecosystem Responses to Climatic Change: Rangelands', in Reddy, K. R. and Hodges, H. F. (eds.), *Climate Change and Global Crop Productivity*, CAB International, N.Y., pp. 293–314.
- Riedo, M., Gyalistras, D., and Fuhrer, J.: 2000, 'Net Primary Production and Carbon Stocks in Differently Managed Grasslands: Simulation of Site-Specific Sensitivity to an Increase in Atmospheric CO<sub>2</sub> and to Climate Change', *Ecol. Modelling* **134**, 207–227.
- Riedo, M., Gyalistras, D., Grub, A., Rosset, M., and Fuhrer, J.: 1997, 'Modelling Grassland Responses to Climate Change and Elevated CO<sub>2</sub>', *Acta Oecologica* **18**, 305–311.
- Rietkerk, M. and van de Koppel, J.: 1997, 'Alternate Stable States and Threshold Effects in Semi-Arid Grazing Systems', *OIKOS* **79**, 69–76.
- Smit, B. and Yunlong, C.: 1996, 'Climate Change and Agriculture in China', *Global Environ. Change* **6**, 205–214.
- Thornley, J. H. M. and Cannell, M. G. R.: 1997, 'Temperate Grassland Responses to Climate Change: An Analysis Using the Hurley Pasture Model', *Ann. Botany* **80**, 205–221.

- Thornley, H. M. and Verberne, E. L. J.: 1989, 'A Model of Nitrogen Flows in Grassland', *Plant Cell Environ.* **12**, 863–886.
- Tucker, C. J.: 1979, 'Red and Photographic Infrared Linear Combinations for Monitoring Vegetation', *Rem. Sens. Env.* **8**, 127–150.
- Ward, S. J. E., Midgley, G. F., Jones, M. H., and Curtis, P. S.: 1999, 'Responses of Wild C4 and C3 Grass (Poaceae) Species to Elevated Atmospheric CO<sub>2</sub> Concentration: A Meta-Analytic Test of Current Theories and Perceptions', *Global Change Biology* **5**, 723–741.
- Watson, R., Noble, I., Bolin, B., Ravindranath, N. H., Verardo, D. J., and Dokken, D. J.: 2000, *IPCC Special Report on Land Use, Land-Use Change, and Forestry*, Cambridge University Press, Cambridge, 377 pp.
- Watson, R. T., Zinyowera, M. C., Moss, R. H., and Dokken, D. J.: 1996, 'Climate Change 1995 – Impacts, Adaptations and Mitigation of Climate Change: Scientific-Technical Analyses', Cambridge University Press, Cambridge, 878 pp.
- Weisberg, P. J. and Coughenour, M. B.: 2003, 'Model-Based Assessment of Aspen Responses to Elk Herbivory in Rocky Mountain National Park, U.S.A.', *J. Environ. Manage.* **32**, 152–169.
- Xiao, X., Ojima, D. S., and Ennis, C. A.: 1997, 'Land Cover Classification of the Xilin River Basin, Inner Mongolia, Using Landsat TM imagery', *Research on Grassland Ecosystems* **5**, 240–252.
- Xiao, X., Shu, J., Yifeng, W., Ojima, D. S., and Bonham, C. D.: 1996, 'Temporal Variation in above Ground Biomass of *Leymus chinense* Steppe from Species to Community Levels in the Xilin River Basin, Inner Mongolia, China', *Vegetatio* **123**, 1–12.
- Yang, H. (ed.): 1987, *Proceedings of the International Symposium on Grassland Vegetation*, Science Press, Hohhot, China, 629 pp.
- Zhai, P., Sun, A., Ren, R., Liu, X., Gao, B., and Zhang, Q.: 1999, 'Changes of Climate Extremes in China', *Clim. Change* **42**, 203–218.
- Zhu, T. C.: 1993, 'Grasslands of China', in Coupland, R. T. (ed.), *Ecosystems of the World. Natural Grasslands: Eastern Hemisphere and Resume*, Elsevier, N.Y., pp. 61–82.

(Received 3 April 2002; in revised form 22 May 2003)

Reproduced with permission of the copyright owner. Further reproduction prohibited without permission.